

Winter CO₂ fluxes in a sub-alpine grassland in relation to snow cover, radiation and temperature

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Abstract Carbon dioxide (CO₂) emissions were measured over a period of 3 years at the sub-alpine Swiss CARBOMONT site Rigi Seebodenalp. Here we show, that winter respiration contributes larger than expected to the annual CO₂ budget at this high altitude, rich in belowground organic carbon grassland (7–15% C by mass). Furthermore the contribution of winter emissions to the annual CO₂ budget is highly dependent on the definition of “winter” itself. Cumulative winter respiration determined over a 6 month period from 15th of October until 15th of April contributed 23.3 ± 2.4 and $6.0 \pm 0.3\%$ to the annual respiration during the years under observation, respectively. The insulation effect of snow and a lowering of the freezing point caused by high concentrations of soil organic solutes prevented the soil from freezing. These conditions favored higher soil temperatures resulting in relatively high respiratory losses. The

duration of snow cover and micrometeorological conditions determining the photosynthetic activity of the vegetation during snow-free periods influenced the size and the variability of the winter CO₂ fluxes. Seasonal values are strongly influenced by the days at the end and the beginning of the defined winter period, caused by large variations in length of periods with air temperatures below freezing. Losses of CO₂ from the snow-covered soil were highest in winter 2003/2004. These high losses were partially explained by higher temperatures in the topsoil, caused by higher air temperatures just before snowfall. Thus, losses are not a consequence of higher soil temperatures registered during the summer heat wave 2003. However, water stress in summer 2003 might have caused an increment in dead organic matter in the soil providing additional substrate for microbial respiration in the following winter. Although considerable day-to-day fluctuations in snow effluxes were recorded, no conclusive and generally valid relationship could be found between CO₂ losses from the snow pack and snow depth, rate of snow melt, wind speed or air pressure. This suggests that time lags and hysteresis effects may be more important for understanding winter respiration than concurrent environmental conditions in most ecosystems of comparable type.

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Introduction

Identifying the terrestrial sources and sinks of carbon dioxide (CO₂) has been the focus of considerable research over the last decades, particularly within global observation networks such as FLUXNET (Baldocchi et al. 2001; Baldocchi 2008). Only a small fraction of these studies, however, have focused on quantifying non-growing season fluxes from seasonally snow-covered systems (Brooks et al. 1997; Gilmanov et al. 2004; Wang et al. 2011; Liptzin et al. 2009), although it is expected that these so-called winter fluxes can contribute a significant component to the annual carbon budget of ecosystems (Brooks et al. 1997; Bubier et al. 2002; Oechel et al. 1997).

Fluxes of CO₂ measured during the growing season provide important information on carbon exchange processes and are well known to be the major contributor for annual C balances. Furthermore, it has been shown for forest ecosystems that CO₂ fluxes are largely influenced by extreme events, e.g. the European heat-wave in 2003 (Ciais et al. 2005). In contrast, such a clear response to the European 2003 heat-wave has not been reported for grassland ecosystems yet (Wohlfahrt 2008) and hence fluxes from outside the growing season may actually be more critically affecting the annual C budgets of grasslands.

There is general consensus that respiratory losses during the cold period may offset the carbon budget measured during the growing season (Aurela et al. 2004; Brooks et al. 2004; Kelley et al. 1968; Lafleur et al. 2001; Oechel et al. 1997; Schimel et al. 2002) in any year, not only in years characterized by extreme events. Cold season respiration rates were already measured when first ecosystem CO₂ flux quantification attempts were started (Johnson and Kelley 1970; Kelley et al. 1968; Vourlitis and Oechel 1999; Zimov et al. 1993). To date, only a limited number of CO₂ flux measurements during the winter have been conducted in northern latitude grasslands, as reviewed by Gilmanov et al. (2004) and recently pointed out in a synthesis study across a variety of ecosystems by Wang et al. (2011). Schimel and Clein (1996) found that wintertime ecosystem respiration in the Western U.S. Mountains can account for carbon losses that total half of the growing season net carbon sequestration, with soil emissions representing a large fraction of this lost carbon. The quantification of winter CO₂

effluxes can thus change both the magnitude and the sign of the annual carbon budget.

CO₂ exchange measurements have especially been made during the last two decades in high latitude and high altitude regions (Merbold et al. 2009; Zeeman et al. 2010; Zhuang et al. 2006), since these ecosystems are particularly vulnerable to climate change (IPCC 2007), contain large soil C stocks (Zimov et al. 2006) and are therefore of special interest for quantifying the global terrestrial CO₂ budget. However, most annual estimates of these ecosystems are still only based on summer CO₂ fluxes (Merbold et al. 2009; Kutzbach et al. 2007; Jones et al. 1998; Oechel et al. 1993).

Environmental conditions, such as rain- and snow-fall, snow cover, air- and soil temperatures, wind speed and air pressure play an important role in winter CO₂ exchange (Bubier et al. 2002; Mariko et al. 2000; Welker et al. 2000) and can vary significantly among years. Bubier et al. (2002) discussed the physical factors and biological processes influenced by snow cover. Heterotrophic respiration in the soil appears to be active even under snow cover, releasing significant amounts of CO₂ (Kelley et al. 1968; Zimov et al. 1993; Oechel et al. 1997; Brooks et al. 2004). Via its function as a thermally isolating ground cover, snow may significantly decrease the depth to which soil freezes, and may even prevent complete freezing in areas without permafrost, thereby increasing the temperature of the active soil layer during winter (Mariko et al. 2000; Swanson et al. 2005). These favorable soil temperatures, with a critical threshold for active respiration around -7 to -5°C , corresponding to the presence or absence of unfrozen water (Brooks et al. 1997; Osterkamp and Romanovsky 1997) allow significant CO₂ production through a continuation of metabolic activity of soil microorganisms with enzymatic systems that are efficient at low temperatures (Panikov and Dedysh 2000). Additionally, root respiration may contribute to the winter CO₂ emissions (Grogan et al. 2001; Welker et al. 2000).

So far no universal relationship between snow pack characteristics and ecosystem respiration has been established. Some authors report on a clear positive effect of snow pack on CO₂ efflux (Gilmanov et al. 2004; Hardy et al. 2001), whereas others were unable to determine an effect of snow cover on respiration (Jones et al. 1999). Of those who identified possible effects of snow cover the processes governing these

effects seemed to be rather specific and not universal. Bubier et al. (2002) found a good correlation between decreasing atmospheric pressure and increasing winter CO₂ efflux; and Kelley et al. (1968) demonstrated a positive effect of wind speed on CO₂ efflux, because accumulated CO₂ within the snow pack is more rapidly released under conditions of high wind speed than under calm conditions. Liptzin et al. (2009) proposed a combination of the previously found relationships, highlighting four different zones of CO₂ emissions driven with increases in, and duration of snow cover: (1) pulses from freeze–thaw processes; (2) CO₂ emissions, which are related to temperature and (3) moisture; and (4) CO₂ emissions influenced by carbon availability. Given this large variability of findings, our study contributes to the overarching goal to increase our mechanistic understanding of how snow cover and snow-related processes govern winter CO₂ fluxes from grassland ecosystems.

Several studies in the arctic tundra calculated CO₂ fluxes using the diffusion gradient of CO₂ concentrations measured at different depths within the snow pack and at the snow–air interface (Brooks et al. 2004; Kelley et al. 1968; Sommerfeld et al. 1996; Swanson et al. 2005). Other studies used automatic chambers (Christensen et al. 2000; Fahnestock et al. 1999; Panikov and Dedysh 2000; Zimov et al. 1993) or the Bowen ratio energy balance technique (Frank and Dugas 2001) to quantify winter CO₂ emissions. Most field studies employed discrete sampling every few days and hence did not collect continuous flux measurements that are necessary to assess the importance of short-term variations of the CO₂ flux from snow-covered ecosystems over the entire cold season. Only very recently a few micrometeorological studies have started to quantify winter net ecosystems exchange using continuous CO₂-flux data (Aurela et al. 2009; Flanagan et al. 2002; Lafleur and Humphreys 2008; Li et al. 2005; Vourlitis and Oechel 1999).

Here we report on continuous eddy covariance (EC) flux measurements carried out during 3 years including the winter periods. Measurements were taken at the Swiss CARBOMONT (Cernusca et al. 2008) site Rigi Seebodenalp, which is snow covered for about 30% of the year. The objectives of this paper are: (1) to quantify the winter CO₂ fluxes measured during the three winter periods; (2) to evaluate the contribution of the winter CO₂ fluxes to the annual carbon budget of

this grassland ecosystem; (3) to investigate the micrometeorological processes controlling winter CO₂ fluxes; and (4) to investigate the temporal evolution of CO₂ fluxes shortly after snow melt.

We hypothesize that the winter derived CO₂ efflux contributes significantly to the site's annual C balance, and that winter respiration is primarily governed by micrometeorological variables such as temperature and wind speed amongst others, since we do not expect a substrate limitation of CO₂ production due to the high organic carbon content in the soil. Additionally, we hypothesize that there must be a continuous loss of CO₂ from the ecosystem to the atmosphere, without distinct peaks at the beginning and the end of the winter season, because soils below the snow cover on grounds without permafrost remain unfrozen all winter.

Methods and site description

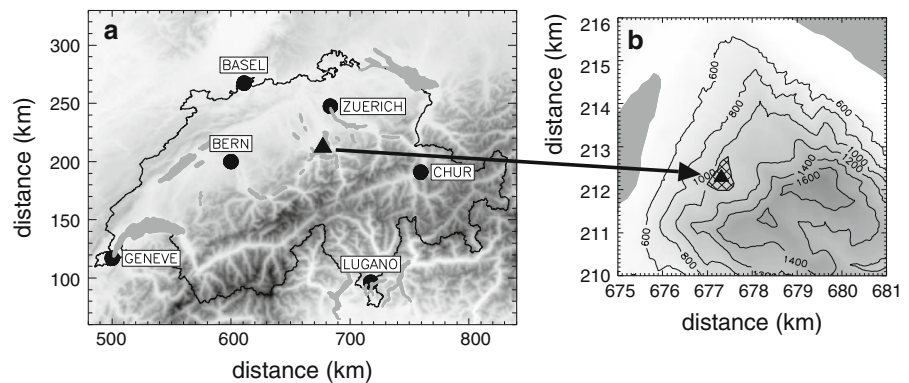
Site description

The Rigi Seebodenalp flux site was established in May 2002 as part of the European Union (EU) funded Framework Program 5 project CARBOMONT. The study area, defined as a sub-alpine grassland, is located on a flat shoulder terrace on the north slope of Mount Rigi (47°03'30.0"N, 8°27'26.5"E) in Central Switzerland at an altitude of 1,025 m above sea level (Rogiers et al. 2005, 2008). The site encompasses 32 ha of relatively flat terrain (Fig. 1). Summer land-use is dominated by extensively grazed pastures and meadows (Rogiers et al. 2005). The current terrain is the bottom of a pre-historic vanished lake, which was fed by melt water at the end of the last glaciation (Vogel and Hantke 1989).

Soil carbon content and vegetation composition

A thick sedge peat layer is found on top of the clay-rich lake sediments. The soils have a very high organic carbon content in the topmost 24 cm ranging between 7.2 ± 0.2 and $15.7 \pm 0.9\%$ by mass and are characterized either as stagnic cambisols or folic (drystic) histosols depending on meso-topographic conditions (Müller 2004; Rogiers et al. 2008). The vegetation type is a *Lolium-*

Fig. 1 Map of Switzerland with indication of the CARBOMONT site Seebodenalp (a). Topography of the northern part of mount Rigi, the position of the Seebodenalp and the measurement tower are shown (b) on a 25-km grid. The maps are in Swiss km-coordinates (DHM25 reproduced with permission, Swisstopo BA046078)



Cynosuretum cristati (rye grass-crested dog's tail) grassland composed of 21–24 vascular plant species with high abundance of *Lolium perenne* L. (rye grass), *Poa pratensis* L. (meadow grass), *Phleum pratense* L. (timothy), *Cerastium holosteoides* Fr. (mouse ear) and *Rumex obtusifolius* L. (bitter dock), but relatively low abundance of *Cynosurus cristatus* L. (crested dog's tail) (Reutlinger 2004). In spring *Ranunculus ficaria* L. (pilewort) and *Anthoxanthum odoratum* L. (scented vernal grass) play an important role (Reutlinger 2004).

Flux measurements

The EC technique was used for continuous measurements of the turbulent fluxes of CO₂, water vapor, sensible heat and momentum (see Aubinet et al. 2000; Baldocchi and Bowling 2003, for methodological details). Wind velocity, wind direction, and temperature fluctuations were measured with a three-dimensional ultrasonic anemometer (Solent HS, Gill Ltd., Lymington, UK), mounted at a height of 2.4 m above ground level (a.g.l.). CO₂ and water vapor concentrations were measured with an open-path infrared gas analyzer (LI-7500, LI-COR Inc., Lincoln, Nebraska, USA). Both instruments were sampled at a resolution of 20 Hz. Data processing and treatment followed CarboEurope standards (Mauder et al. 2008) with site-specific details described in Rogiers et al. (2008). In brief, the vertical turbulent fluxes (F) were calculated from the half-hourly averaged covariance of the measured fluctuations of the vertical wind velocity (w , m s⁻¹) in a coordinate system which was aligned with the mean streamlines (McMillen 1988) and the CO₂ concentration (c , μmol):

$$F = \overline{w'c'} \quad (\mu\text{mol m}^{-2} \text{s}^{-1}) \quad (1)$$

Overbars denote time averages and primed quantities are the instantaneous deviations from their respective time average. The linear trend in CO₂ concentration was subtracted from each half hour interval to obtain c' . CO₂ fluxes were corrected for high-frequency damping losses (Eugster and Senn 1995) using a damping constant of 0.2 s⁻¹, followed by the necessary density flux correction for open-path instruments according to Webb et al. (1980) and a correction for self-heating of the instrument (the “Burba” correction in a variant suitable for the Rigi Seebodenalp site as described by Rogiers et al. 2008).

Micrometeorological data

To characterize climatic conditions additional meteorological sensors were installed near the EC tower on a separate tower. Air and soil temperatures (°C) were measured with copper-constantan thermocouples at the heights 100, 50, 10, 5 cm aboveground and at −5, −10, −20, −30, −50 cm belowground, measured from the soil surface. Soil moisture data were available for the growing season only, caused by regular sensor failure. Wind speed was measured with a switching anemometer (Vector Instruments, UK) at 2 m height (m s⁻¹). All data were recorded with a data logger (model CR10X, Campbell Scientific Inc., Loughborough, UK).

Snowfall and snow depth were measured manually on a daily basis using a graduated measuring rod at the nearby station Oberiberg at similar altitude (1,090 m a.s.l.) and exposition (north), 6.2 km southeast from Rigi Seebodenalp. The quantitative Oberiberg snow data correspond well with the qualitative information

on snow cover provided by local people from Seebodenalp.

Data availability and post-processing

The EC tower and the micrometeorological station were continuously operated from 17th May 2002 until 10th May 2005, including the three winter periods. Due to some technical failures of the EC system, the data availability before filtering amounted to 88% of all possible 30-min time periods. The data were screened for outliers based on a set of testable plausibility criteria as described in Rogiers et al. (2005). This led to rejection of 7% of available data due to snow- and rainfall and 32% of the data were discarded due to inadequate turbulence conditions, which occurred mainly (65% of these cases) at night (photosynthetic photon flux density (*PPFD*) < 10 $\mu\text{mol m}^{-2} \text{s}^{-1}$). The coverage of high quality data during the three winter periods thus was 61%.

Calculations

CO₂ fluxes were calculated as 30-min averages (Eq. 1) for a 3-year period from May 2002 to May 2005. The annual CO₂ integrals for the 3 years of measurement were obtained by integrating the CO₂ flux data from 15th May until 14th May of the following year. The winter period—or cold season—in this paper was defined as the period from 15th October until 15th April, covering all days when snow blanketed the site in any of the 3 years. This period is longer than the conventional definition of winter but essential due to the long time interval when the study site is covered with snow. The only exceptions were very few anomalous single-day snow events in autumn and spring. It is to be noted that in the Alps at this altitude snow events can occasionally even happen during the warm season. Therefore, our choice of the definition of winter differs from the arbitrary definition of winter in climatology (December–February), but it was not possible to restrict our analysis to December–February, since this would have left too many days with snow cover in the wrong season.

Net ecosystem exchange (*NEE*) was partitioned into gross primary production (*GPP*) and ecosystem respiration (*Reco*). For days when the vegetation was photosynthetically active (i.e. days where there was a

clear response of *NEE* to *PPFD*), total ecosystem respiration was calculated as the weighted average of nighttime respiration (*Reco_n*) and daytime respiration (*Reco_d*). Nighttime (*PPFD* < 10 $\mu\text{mol m}^{-2} \text{s}^{-1}$) CO₂ fluxes are assumed to represent ecosystem respiration at night. Daytime respiration was derived from the light response curves calculated for each day when the vegetation was photosynthetically active. The light response curve defines the relationship between CO₂ exchange during the day (*NEE*) and *PPFD* (*PPFD* > 10 $\mu\text{mol m}^{-2} \text{s}^{-1}$), which can be described by a rectangular hyperbola (Gilmanov et al. 2003; Ruimy et al. 1995):

$$NEE = [(-F_{\infty} \times \alpha \times PPFD) / (\alpha \times PPFD + F_{\infty})] + Reco_n \quad (2)$$

where F_{∞} is *NEE* at light saturation ($\mu\text{mol m}^{-2} \text{s}^{-1}$), α is the apparent quantum yield, and *Reco_n* ($\mu\text{mol m}^{-2} \text{s}^{-1}$) is to be interpreted as the best estimate of the average *Reco_d* (Gilmanov et al. 2003; Suyker and Verma 2001).

GPP was then calculated as the difference between *NEE* and *Reco*. Therefore, negative values indicate CO₂ uptake by photosynthesis and positive values the release of CO₂ by respiration. For days without a clear response of *NEE* to *PPFD*, total ecosystem respiration was calculated from the exponential relationship between *Reco_n* and soil temperature at –5 cm (°C) (Schmid et al. 2000),

$$Reco_n = a^{(b \times T_s)}, \quad (3)$$

where a and b are parameters determined by minimizing the sum of squares of the residuals and T_s (°C) represents the soil temperature at –5 cm.

The light-response approach was also used to fill gaps in our data to obtain annual and seasonal totals. Outside the growing season, however, fluxes did not show a clear dependency on *PPFD* due to the lack of photosynthetic activity. Hence, gaps in the data had to be filled differently for the winter period: short gaps (≤ 2 h) were filled by linear interpolation, while larger gaps were filled using the median flux value for each hour of the day, determined from available data 3 days prior and after the gap. Gap-filled data only served the purpose to obtain annual and seasonal budgets, but were not used to find relationships between snow cover and CO₂ flux.

Results

Snow cover

Snow coverage at Rigi Seebodenalp varied greatly among winters (Table 1; Fig. 2). During the years 2002/2003, 2003/2004 and 2004/2005 the site was snow covered during 88, 125 and 116 days. On average, Rigi Seebodenalp was snow covered during 30% of the year. Also the timing of the first (autumn) and the last (spring) day with snow cover was very variable during the three winter periods (Fig. 2). First snow in 2004 [day of year (DOY) 313; 9th November] blanketed the site 47 days later than in 2002 (DOY 266). In spring, the last snow melted much earlier in 2003 (DOY 102) compared to 2004 (DOY 130). Maximum snow depth was 70 cm.

Winter CO₂ fluxes

Daily CO₂ fluxes during the winter period (15th October–15th April) ranged from a net CO₂ uptake of $-15.17 \text{ g C m}^{-2} \text{ day}^{-1}$ in spring 2005 (DOY 105) to a net CO₂ loss of $12.39 \text{ g C m}^{-2} \text{ day}^{-1}$ in February 2004 (DOY 51). The mean daily CO₂ fluxes (F ; Table 1) varied substantially over the three winter periods, with moderate fluxes during 2002/2003 ($2.09 \pm 0.21 \text{ g C m}^{-2} \text{ day}^{-1}$), largest values during 2003/2004 ($3.39 \pm 0.19 \text{ g C m}^{-2} \text{ day}^{-1}$) and smallest fluxes during 2004/2005 ($1.27 \pm 0.25 \text{ g C m}^{-2} \text{ day}^{-1}$).

A similar picture was given, when focusing on snow-covered days only, mean CO₂ losses (F_{snow} ; Table 1) were remarkably higher in 2003/2004 ($4.33 \pm 0.18 \text{ g C m}^{-2} \text{ day}^{-1}$) than during the two other winter periods ($3.03 \pm 0.31 \text{ g C m}^{-2} \text{ day}^{-1}$ in 2002/2003; $2.63 \pm 0.17 \text{ g C m}^{-2} \text{ day}^{-1}$ in 2004/2005). This, together with the greater number of snow covered days (D_{ts} ; Table 1) in 2003/2004 (125 days) contributed substantially to the observation that the winter 2003/2004 had the highest mean daily CO₂ losses.

During snow free days in the winter period, mean daily net CO₂ fluxes fluctuated around zero with some days showing a net uptake but others showing a net carbon loss (Fig. 2). We distinguished between snow free days when there was no photosynthetic activity and thus daily gross primary production (GPP_d) $\geq -0.5 \text{ g C m}^{-2} \text{ s}^{-1}$ (F_{nsna} ; Table 1) and

Table 1 Daily averages measured at Seebodenalp during three winter periods (183 days) from 15 October (day 288) until 15 April (day 105) in 2002–2003, 2003–2004 and 2004–2005 of soil temperature under snow cover at 5 and 30 cm depth for January–March (T_{s5} , T_{s30}), CO₂ flux over snow pack (F_{snow}), CO₂ flux without snow cover when GPP at noon $\geq -0.5 \text{ g C m}^{-2}$ (F_{nsna}) and CO₂ flux without snow cover when GPP at noon $< -0.5 \text{ g C m}^{-2}$ (F_{nsd})

Period	T_{s5} (°C)	T_{s30} (°C)	F ($\text{g C m}^{-2} \text{ day}^{-1}$)	F_{snow} ($\text{g C m}^{-2} \text{ day}^{-1}$)	F_{nsna} ($\text{g C m}^{-2} \text{ day}^{-1}$)	F_{nsd} ($\text{g C m}^{-2} \text{ day}^{-1}$)	D_{ts} (days)	D_{f} (day)	D_{l} (day)	D_{gd} (days)
2002/2003	0.88 ± 0.02	1.65 ± 0.08	2.09 ± 0.21	3.03 ± 0.31	3.32 ± 0.22	1.35 ± 0.22	73	88	266	102
2003/2004	1.20 ± 0.07	1.61 ± 0.09	3.39 ± 0.19	4.33 ± 0.18	3.22 ± 0.20	1.80 ± 0.22	111	125	278	130
2004/2005	0.41 ± 0.02	0.74 ± 0.05	1.27 ± 0.25	2.63 ± 0.17	1.75 ± 0.24	-0.78 ± 0.32	93	116	313	104
										58

D_{ts} is the number of days with snow coverage within the winter period, D_{f} is the number of total snow covered days for the three measurement years, D_{l} and D_{gd} are the days on which first (autumn) and last (spring) snow, respectively, occurred. D_{gd} is the number of growing days (mean air temperature $> 5^\circ\text{C}$)

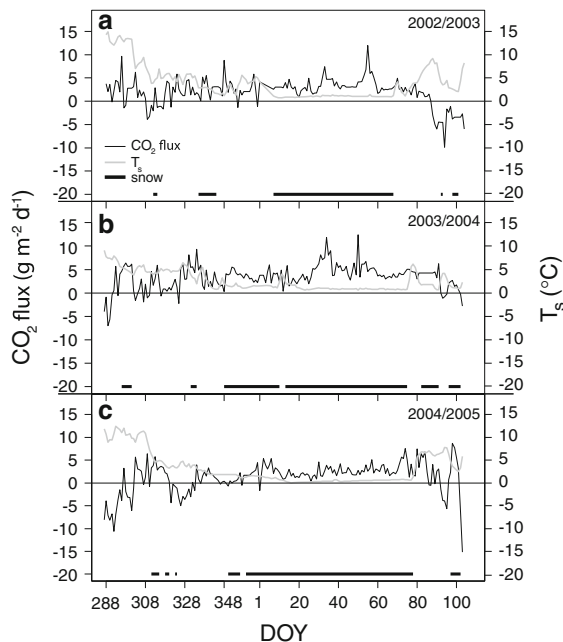


Fig. 2 Mean daily CO₂ fluxes ($\text{g C m}^{-2} \text{ day}^{-1}$), mean soil temperature at -5 cm ($^{\circ}\text{C}$, T_s) and snow cover presence (snow) during the three winter periods 2002/2003 (a), 2003/2004 (b) and 2004/2005 (c). Winter periods start at 15th October (day = 288) until 15th April (day = 105)

snow-free days where assimilation occurred and $\text{GPP}_d < -0.5 \text{ g C m}^{-2} \text{ s}^{-1}$ (F_{nsa} ; Table 1). CO₂ fluxes during snow free days without daily photosynthetic activity $\geq -0.5 \text{ g C m}^{-2} \text{ s}^{-1}$ (F_{nsda}) varied among the three winter periods between 1.75 ± 0.24 and $3.32 \pm 0.22 \text{ g C m}^{-2} \text{ day}^{-1}$ (Table 1). Mean daily CO₂ exchange for snow-free days with photosynthetic activity (F_{nsa}) varied substantially. Net daily mean losses were measured in winter 2002/2003 ($1.35 \pm 0.22 \text{ g C m}^{-2} \text{ day}^{-1}$) and in winter 2003/2004 ($1.80 \pm 0.22 \text{ g C m}^{-2} \text{ day}^{-1}$). Contrastingly, a daily mean uptake was recorded in winter 2004/2005 ($-0.78 \pm 0.32 \text{ g C m}^{-2} \text{ day}^{-1}$).

Annual CO₂ budgets, contribution of NEE and snow covered days

The interannual variability in CO₂ budgets was very high for the three winter periods (Fig. 3) and the annual CO₂ integrals (NEE_a ; Table 2) demonstrate that the site was a net source of C during all years under observation. Over the winter period (NEE_w ; Table 2) 2003/2004 the highest carbon loss was

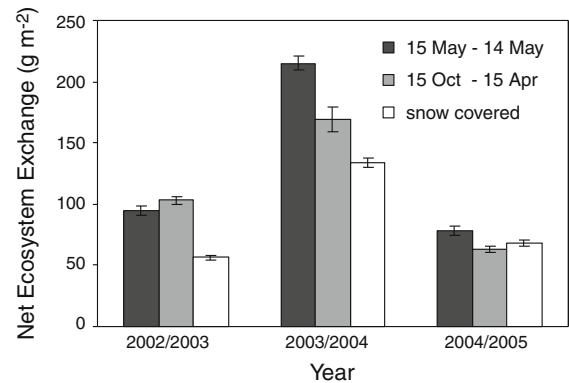


Fig. 3 NEE budgets (g C m^{-2}) at Rigi Seebodenalp, where a positive sign means a net loss. The annual NEE budget (calculated from 15th May until 14th May in the following year), the winter NEE budgets (calculated from 15th October until 15th April) and the NEE budget during snow covered periods were calculated for the three measurement years 2002/2003, 2003/2004 and 2004/2005

observed ($169 \pm 4 \text{ g C m}^{-2}$). Significantly lower losses were registered during the other winters ($103 \pm 3 \text{ g C m}^{-2}$ in 2002/2003 and $63 \pm 2 \text{ g C m}^{-2}$ in 2004/2005). Still, the reported values show that winter emissions have a severe impact on the annual CO₂ budget. The share of carbon losses during snow covered days (NEE_s ; Table 2) in 2002/2003 and 2004/2005 amounted to 56 ± 2 and $68 \pm 2 \text{ g C m}^{-2}$, respectively. CO₂ efflux in 2003/2004 after the summer heat-wave in Europe (see Schär et al. 2004; Ciais et al. 2005) was twice as high ($133 \pm 4 \text{ g C m}^{-2}$).

When partitioning NEE into respiration and assimilation, it becomes clear that the highest contribution of snow covered periods ($Reco_s$; Table 2) to the annual ecosystem respiration was measured in 2003/2004. During these 2 years, $Reco_s$ contributed 6.5% to the total annual respiration ($Reco_a$; Table 2). The share of $Reco_s$ to $Reco_a$ was slightly lower in 2002/2003 (5.7%) and 2004/2005 (5.9%).

Highest NEE values measured during the winter period 2003/2004 (F ; Table 1) are not the result of especially high winter respiration rates ($Reco_w$; Table 2), but are rather the result of intermediate respiration rates over the whole winter in combination with relatively low assimilation rates (GPP_w ; Table 2) and the lowest number of growing degree days ($Dgd = 18$; Table 1). The share of ecosystem respiration measured during winter ($Reco_w$; Table 2) to the

Table 2 CO₂ budgets (g C m⁻²) for the 3 years of measurements at Seebodenalp of net ecosystem exchange of CO₂ (*NEE*), of ecosystem respiration (*Reco*) and of gross primary production (*GPP*)

Year	<i>NEE_a</i>	<i>NEE_w</i>	<i>NEE_s</i>	<i>Reco_a</i>	<i>Reco_w</i>	<i>Reco_s</i>	<i>GPP_a</i>	<i>GPP_w</i>	<i>GPP_s</i>	<i>Reco_{w3}</i>
2002/2003	95 ± 4	103 ± 3	56 ± 2	1337 ± 19	374 ± 7	76 ± 3	-1242 ± 15	-274 ± 5	-15 ± 1	154 ± 5
2003/2004	215 ± 6	169 ± 4	133 ± 4	1602 ± 24	347 ± 7	140 ± 4	-1395 ± 18	-184 ± 3	-11 ± 2	150 ± 6
2004/2005	78 ± 5	63 ± 2	68 ± 2	1331 ± 20	269 ± 5	79 ± 2	-1256 ± 16	-208 ± 3	-13 ± 1	88 ± 4

The total annual budgets (suffix *a*), the budgets of CO₂ fluxes measured during winter from 15 October until 15 April (suffix *w*) and the budgets during snow cover (suffix *s*) are listed. *Reco_{w3}* is the respiration budget for the 3 months winter period from 1st of December until 28th of February

total annual respiration (*Reco_a*; Table 2) varied between 28% (in 2002/2003) and 20% (in 2004/2005).

There was some photosynthetic activity during snow covered days that led to *GPP_s* < 0 (Table 2). Days with measurable assimilation only occurred towards the end of the snow period where both, snow-free and snow-covered patches were found within the footprint of the EC tower.

Environmental variables and CO₂ fluxes

In winter, air temperatures regularly dropped below 0°C. However, the soil at Rigi Seebodenalp never froze down to -5 cm (topsoil), where our near-surface soil temperature sensor was located, not even during snow free periods. Air-, snow-, and soil temperatures for a period with 40 cm snow cover in February–March 2005 were analyzed in detail (Fig. 4). Air temperatures at 2 m height were constantly below freezing. 10 cm above the snow (at 50 cm height) and in the snow pack (10 and 5 cm) temperatures were less

negative and their peak values were slightly lagging air temperature. Temperature fluctuations in the air were strongly damped by the snow pack and were thus smallest near the soil. Below the snow cover, soil temperatures increased with depth (Fig. 5). The temperature of the first 50 cm of the soil under the snow pack (*T_{s5}*; Table 1) was higher in January–February 2004 than during the same months in the winter periods 2002/2003 and 2004/2005. Temperature in the uppermost soil layer (-5 cm) below the snow pack fluctuated between 0 and 2°C, but never dropped below 0°C.

In summer 2003, Europe experienced extremely hot temperatures (Schär and Jendritzky 2004; Luterbacher et al. 2004; Ciais et al. 2005). These high air temperatures in summer 2003 also translated into higher summer soil temperatures. However, the influence of this extreme event did not persist until the following winter. A cold snap in autumn 2003 (2nd October; day 273) reduced soil temperatures

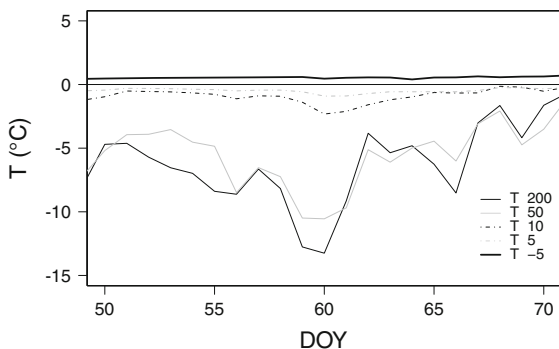


Fig. 4 Air (200, 50 cm), snow (10, 5 cm) and soil temperatures (-5 cm) for a period with 40 cm snow cover in February–March 2005 (days 50–70). Temperature fluctuations were damped from the free atmosphere over the snow towards the soil. The 0°C line is added to the graph

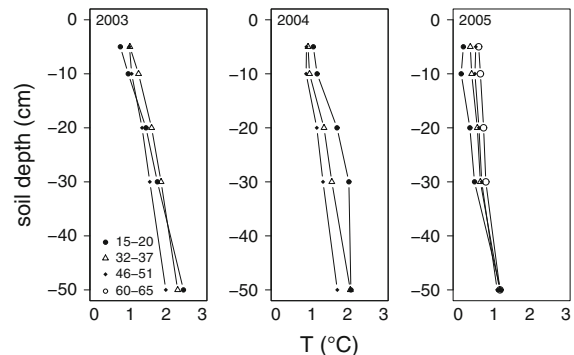


Fig. 5 Soil temperature profiles (°C) for four time intervals in spring 2002, 2003 and 2004. The four intervals are snow covered periods within the period 20 January (day 20)–10 March (day 69). Before day 20, there was at least 5 days of continuous snow cover in all winters

considerably, resulting in lower mean values in autumn 2003 compared to 2002 and 2004.

Since the temperature of the uppermost soil layer under snow cover fluctuated in the range 0–2°C, we hypothesized a constant CO₂ efflux from snow cover. However, considerable fluctuations in respiratory losses were measured over the snow pack (Fig. 2). We investigated the relationship between respiratory losses from snow covered grassland and meteorological variables such as snow depth, rate of snow melt, wind speed (Fig. 6), soil temperature (not shown) and air pressure (not shown). None of these variables explained more than 10% of the total variance in CO₂ efflux during periods with snow cover.

A more significant relationship was however found between snowmelt rates and CO₂ fluxes whenever snow had lasted for a duration of at least 7 days (Fig. 7). The diurnal patterns of the CO₂ flux (*NEE*) and soil temperature (*T_s*) are best illustrated by the interquartile range (the range between the 75%-quantile and the 25%-quantile of the 30-min fluxes of each day), which we denote as ΔNEE and ΔT_s , respectively (Fig. 7a–c). As long as the vegetation is inactive, we expect the interquartile range of *NEE* to be close to zero, thus no relevant diurnal course is expected. Similarly, ΔT_s is expected to be close to zero whenever the ground is snow covered or frozen, due to the fact that the phase change when melting or freezing water at temperatures close to 0°C consumes or produces heat and thereby reducing the daily amplitude of soil temperature fluctuations (Fig. 7, days before day 0). To separate conditions with

inactive vegetation from such where it should be active, we used threshold values of 1.5 $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$ for ΔNEE and 0.5 K for ΔT_s .

As soon as the snow pack had disappeared, *T_s* immediately increased and a diurnal pattern ($\Delta T_s > 0.5 \text{ K}$) was observed (Fig. 7a–c, days after day 0). The increase in *T_s* was correlated with the increase in photosynthetic activity, which was reflected in ΔNEE and in a decrease in *GPP*. However, the start of assimilation (*GPP*, negative values) and of the measured diurnal pattern of *NEE* did not coincide precisely with the time of snowmelt or the start of a measurable diurnal cycle in soil temperature (Fig. 7a, c). In 2004, photosynthetic activity was even observed 3 days before the diurnal pattern in soil temperature was detectable, which can be explained by the spatial heterogeneity of the snow pack during the period of snow melt (Fig. 7b).

In spring 2004, a clear increase in soil temperature was observed after snow had disappeared, and a net CO₂ uptake (*NEE* < 0) was measured from the very first day after snow melt (Fig. 7b). In 2003, it took 19 days from the time when snow had melted near the tower until a daily mean uptake was registered. Immediately after snow melt, changes in mean daily *NEE* were less pronounced than changes in ΔT_s .

Photosynthetic activity in spring

The photosynthetic rates in spring after day 104 (i.e. day of snowmelt in 2004) were similar during

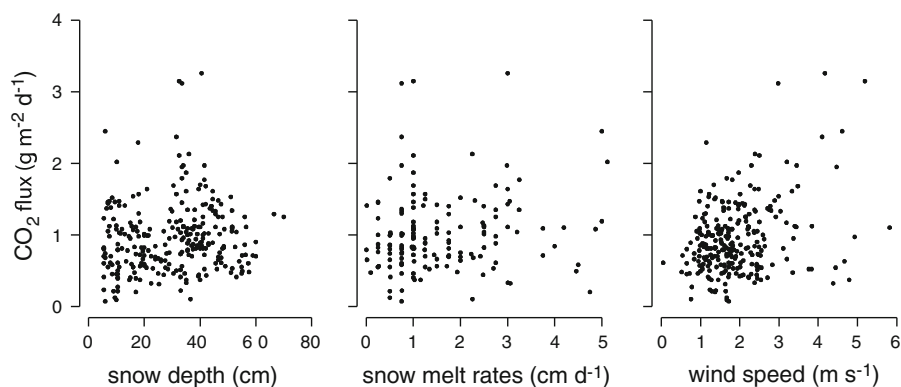
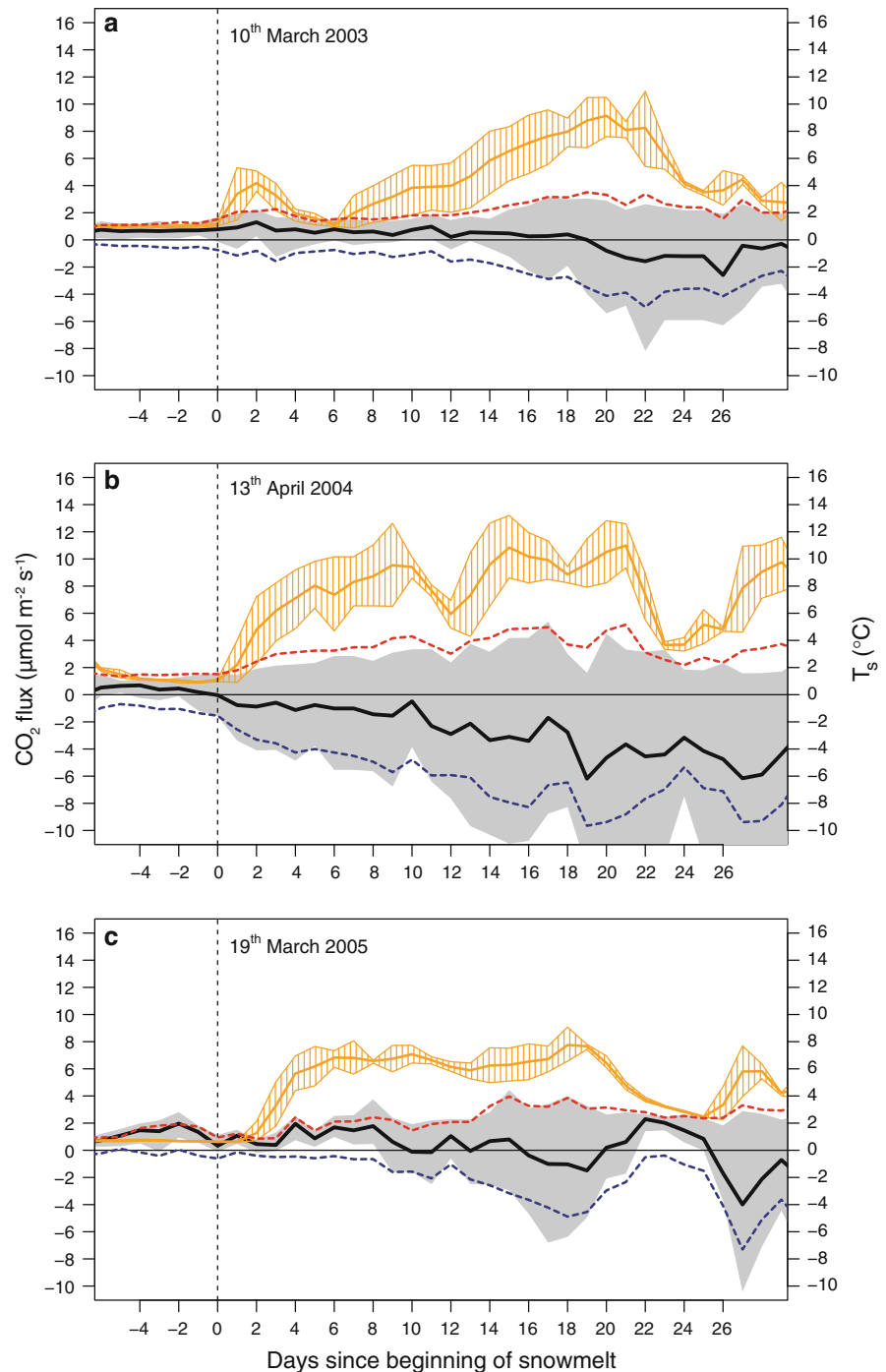


Fig. 6 Relationship between net daily CO₂ efflux ($\text{g C m}^{-2} \text{ day}^{-1}$) from snow pack and snow depth, snow melt rates, and wind speed for snow covered days at Seebodenalp. None of the relationships explained more than 10% of the

variance and we could thus not determine any statistically significant relationship between CO₂ efflux and the micrometeorological variables

Fig. 7 Daily means of net CO_2 flux (NEE ; $\mu\text{mol CO}_2 \text{ m}^{-2} \text{ s}^{-1}$), ecosystem respiration ($Reco$), gross primary production (GPP) and soil temperature (T_s ; $^{\circ}\text{C}$) at the beginning of spring (complete snow melt at day = 0) for the three winter periods 2002/2003, 2003/2004, and 2004/2005. The diurnal patterns of NEE (black solid line) and T_s (orange solid line) are visualized by the polygon areas covering the range between the 25 and 75% quantiles of NEE and T_s



the 3 years (Fig. 8). After maximum carbon losses in spring and before the first grass cut at the beginning of June, the slopes of the integrated assimilation fluxes (GPP) had similar steepness,

indicating that the increment in net daily CO_2 flux is rather constant for the 3 years and hence largely unrelated to snow conditions of the previous winter.

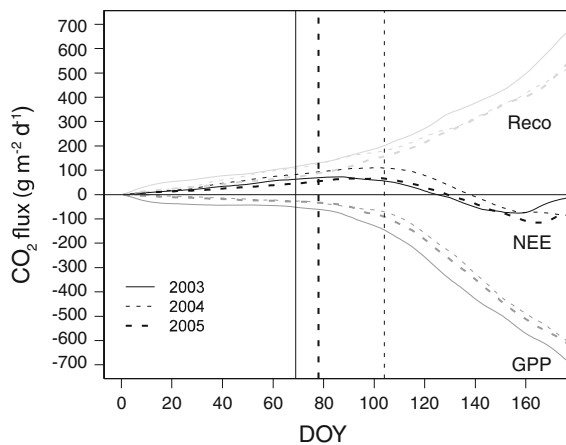


Fig. 8 Cumulative carbon fluxes of net CO_2 flux (NEE), of the respiration ($Reco$) and the assimilation (GPP) fluxes at Rigi Seebodenalp in spring 2003, 2004, and 2005. CO_2 integrals were calculated by integrating the CO_2 flux data from 1 January (day = 1). A positive sign means a net C loss. The vertical lines indicate the timing of the snow melt for each year

Discussion

Magnitude of CO_2 fluxes measured at Rigi Seebodenalp

The mean daily winter CO_2 efflux rates measured at Rigi Seebodenalp (Table 1) are situated in the upper range of available studies on respiration losses from seasonally snow covered non-forest ecosystems. Ham and Knapp (1998) reported average rates of winter-time CO_2 efflux of $0.95 \text{ g C m}^{-2} \text{ day}^{-1}$ measured at a tallgrass prairie and Volk and Niklaus (2002) detected respiratory losses of $1.2 \text{ g C m}^{-2} \text{ day}^{-1}$. The high soil organic matter content at Rigi Seebodenalp (between 7.2 ± 0.2 and $15.7 \pm 0.9\%$ by mass) is certainly one of the relevant factors responsible for the high winter CO_2 efflux rates reported in this study. Average winter CO_2 respiration rates recorded in the arctic tundra are substantially lower than the values measured at Rigi Seebodenalp. Values for the arctic tundra are reported between $0.06 \text{ g C m}^{-2} \text{ day}^{-1}$ (Fahnestock et al. 1999) and $0.18 \text{ g C m}^{-2} \text{ day}^{-1}$ (Bubier et al. 2002). It appears that although arctic and alpine vegetation types are often considered to be similar, the presence (arctic) or absence of permafrost (below 2,580 m a.s.l. in the Alps; see Luetsch et al. 2004, 2004) are distinct differences for winter respiration rates.

Environmental variables and winter CO_2 fluxes

Winter CO_2 efflux is largely the result of microbial respiration, which is sensitive to soil temperature and soil moisture availability (Edwards and Cresser 1992; Liptzin et al. 2009; Schadt et al. 2003). Microbial activity increases exponentially with warmer soil temperatures (Mikan et al. 2002). The higher respiration rates under snow cover (F_s ; Table 1) in 2003/2004 compared to the other two winters can thus be interpreted as the result of the higher soil temperatures (T_{ss} ; Table 1). Deeper in the soil, the temperature was not significantly different between spring 2003 and spring 2004 due to cooling of this layer because of the constant soil heat flux towards the upper soil layer.

Liquid water, which is a prerequisite for cellular activity (Jones et al. 1999; Mikan et al. 2002), was available at Rigi Seebodenalp during winter and was favored by positive soil temperatures. The soil at Rigi Seebodenalp was never frozen down to -5 cm , neither during snow covered nor during snow free periods since snow cover effectively decouples soil temperatures from the atmosphere (Brooks et al. 2004) and keeps soils from deep-freezing due to insulation (Haebleri 1973; Hardy et al. 2001; Walker et al. 1999; Bubier et al. 2002; Shibistova et al. 2002; Swanson et al. 2005). But also during snow-free days with atmospheric frost, the soil at Rigi Seebodenalp did not freeze except for the topmost centimeters ($\leq 5 \text{ cm}$) above our uppermost soil temperature sensor. Brooks et al. (1997) observed that in alpine tundra organic solutes act to suppress the freezing point in the organic soil horizon. The same observation was already described by Edwards and Cresser (1992), who found that a solute depression of the freezing point by ion diffusion as soils begin to freeze may result in thawed soils at temperatures below 0°C , which in combination with the geothermal heat flux due to the absence of permafrost is most likely responsible for the productive environmental conditions that the soil microorganisms experience in winter. Our measurements of soil temperature in the upper 50 cm of the soil indicate that heat, stored in the ground, is released in winter to keep snow basal temperatures close to 0°C (Haebleri 1973). This relation is only valid for systems without underlying permafrost. The temperature of the uppermost soil layer (-5 cm) below the snow pack fluctuated between 0 and 2°C , but never dropped below 0°C , as observed in mountainous areas

(Haeblerli 1973). This does not mean that it is impossible that temperatures can drop below 0°C, and Sevanto et al. (2006) reported a short period where this general rule did not apply, probably caused by extremely dry and porous ground where cold air advection may have been responsible for temperatures as low as −4°C at 5 cm depth. However after roughly 3 weeks of snow cover this short-term deviation from the rule converged to the generally expected 0 to −1°C constant temperature in the mineral soil layer (Sevanto et al. 2006, their Fig. 1).

We were unable to determine any significant relationship between CO₂ efflux from the snowpack and micrometeorological variables such as soil temperature, snow depth, rate of snow melt, wind speed or air pressure. In the special case of soil temperature, the observed range of values under the snowpack was just too narrow (0–2°C) to describe CO₂ flux over snow pack as a function of soil temperature. This is in agreement with several other studies, which also did not find good correlations between the two variables (Nadelhoffer et al. 1991; Sommerfeld et al. 1993; Winston et al. 1997; Schmidt et al. 1999; Suni et al. 2003), although they were based on much wider soil temperature ranges. For tussock–tundra soils, Schimel and Clein (1996) found that there is generally very little response of soil respiration with temperatures below 10°C. Laboratory studies (Nadelhoffer et al. 1991) however indicate that there should be a strong relationship between soil temperatures and respiration even in frozen soils down to −10°C. Our data do not contradict this finding, but they clearly show that the observed range in soil temperatures under natural conditions do not allow to establish such a relationship.

Additionally, we could not find a clear effect of snow depth on CO₂ efflux, which is in agreement with the findings by Jones et al. (1999). In contrast, other authors (e.g. Welker et al. 2000; Gilmanov et al. 2004) determined a positive effect of snow depth on CO₂ efflux. Gilmanov et al. (2004), summarizes two opposing effects of snow cover on CO₂ exchange found in the literature: (1) depending on snow thickness and snow properties, snow cover can act as an almost impermeable coat and can thus significantly decrease the efflux of CO₂ to the atmosphere (Kelley et al. 1968); but (2) a thicker snow cover insulates the soil better and can create a more favorable

environment for belowground respiration during winter (Welker et al. 2000; Liptzin et al. 2009).

Winter CO₂ budgets

The marginal CO₂ uptake in the winter of 2004/2005 could be explained by an extended autumn with photosynthetic uptake, and an earlier start of the growing season. In combination, this partially explained the relatively small daily mean loss determined over the whole winter period 2004/2005. The other important component that led to these small losses was that the winter period 2004/2005 was characterized by the highest number of growing degree days as similarly shown in a synthesis paper by Wang et al. (2011).

During the snow covered periods in 2003/2004 higher carbon losses were measured compared to the other two winter periods (Tables 1, 2). The higher temperatures of the topmost soil (T_{s5} ; Table 1) are partially responsible for these differences (e.g. Reichstein et al. 2003, 2005). Another possible explanation for the high carbon losses measured during the winter following the summer heat-wave 2003 could be the flush of decomposition occurring after rewetting a dry soil (Franzluebbers et al. 2000). During the hot summer 2003 there were some periods where the vegetation at Rigi Seebodenalp suffered from water shortage. Due to water stress, a part of the soil microbial biomass and the roots died and this material is respired when the soil is rewetted. The process of drying and rewetting the soil generally results in an increment of the amount of available substrate, which forms the feeding basis for microbial respiration.

However, we could not detect the CO₂ flush following snow melt/rewetting as proposed by Franzluebbers et al. (2000, Figs. 7, 8), from the melting snow pack nor from the thawing of the soil or a larger availability of nutrient, which clearly differs from previous findings described in literature (e.g. Skogland et al. 1988; Panikov and Dedysh 2000; Prieme and Christensen 2001; Bubier et al. 2002). At Rigi Seebodenalp, there were several snow-melting events during the three winter periods (Fig. 2), the soil was never frozen and thus no freeze–thaw cycles were observed as reported in Liptzin et al. (2009). From this we conclude that we have a unique dataset that illustrates how environmental conditions in the Alps

differ from published knowledge with respect to winter CO₂ effluxes.

Definition of winter

As noted before, we chose a special definition for “winter” to match the true pattern of snow cover as well as possible. In this paper, we defined the winter season as the period from 15th October until 15th April. If we were to use another definition for winter, e.g. only the calendar months December, January, and February, winter CO₂ losses would only account for 11.5, 9.5, and 6.5% of the respective annual CO₂ budgets during the three years under observation (May–April). This definition would lead to a serious underestimation of the importance of winter conditions (characterized by snow cover at our location) in the annual balance. Irrespective of the definition of winter, we found that the snow and frost-free days at both the beginning and ending of the cold season can contribute considerably to the high interannual variability of mean daily fluxes and thus of the winter and growing season carbon budgets. Other studies also had to adopt the definition of winter to their locality (Oechel et al. 1997; Mast et al. 1998; Hobbie et al. 2000; Panikov and Dedysh 2000; Lafleur et al. 2001; Wickland et al. 2001; Hirano 2005; Wang et al. 2011). Therefore, a comparison with literature values not only depends on ecosystem properties, but also on the length of the winter season and micrometeorological conditions. The reason for this is related to the general issue of extensive versus intensive variables (see Legendre and Legendre 1998): winter respiration is an extensive variable “whose values, in a homogeneous system, changes proportionally (linear relationship) to the size of the sampling unit (transect, quadrat, or volume [in our case: length of time period])” (Legendre and Legendre 1998). Comparable estimates of winter respiratory losses fluctuate between 3 and 50% of the annual respiratory CO₂ losses for northern wetlands, arctic tundra (Oechel et al. 1997; Hobbie et al. 2000; Panikov and Dedysh 2000; Lafleur et al. 2001; Hirano 2005) and 10–20% in alpine and subalpine ecosystems (Mast et al. 1998; Wickland et al. 2001).

A special assessment on the definition of winter and how the corresponding values were affected, was published by Bubier et al. (2002). They demonstrated that the contribution of winter respiration to the total

annual respiration varied between 3 and 25%, depending on the winter period definition.

Conclusions

Substantial amounts of CO₂ losses were observed at the sub-alpine Swiss CARBOMONT site Rigi Seebodenalp during three winter periods (integrated from 15th October until 15th April). Total winter respiration and respiration from snow pack contributed 23.3 ± 2.4 and $6.0 \pm 0.3\%$, respectively, to the annual respiration losses at our research site. These winter carbon losses account for an important share in the annual CO₂ budgets, which emphasizes the importance of quantifying CO₂ fluxes outside the growing season.

Mean daily CO₂ fluxes in winter ranged from a net uptake to a net loss. The variability in winter fluxes was strongly determined by the CO₂ losses from snow pack, and by the micrometeorological conditions whenever there was no snow. Especially the days at the beginning and the end of the cold season had a strong influence on the seasonal budget, which therefore should be included in the definition of “winter” at our and comparable locations. The highest daily mean losses were recorded in winter 2003/2004. This is the combined result of: (1) the high respiration from the snowpack ($F_s = 4.33 \pm 0.18 \text{ g C m}^{-2} \text{ day}^{-1}$); (2) the long persistence of snow cover (111 days); and (3) the relatively low photosynthetic activity ($GPP_w = 184 \pm 3 \text{ g C m}^{-2}$). Although the winter 2003/2004, which followed the European summer heat-wave 2003 had twice the respiration of the two other winters, we did not find sufficient evidence to attribute this to the special conditions observed during the heat-wave. This may be a result of the functional difference of grassland ecosystems when compared to forest ecosystems (Ciais et al. 2005) because of the much shorter C turnover time in grasslands as compared to forests, a fact needing further scientific attention. CO₂ exchange over grasslands in Europe has not been investigated as long as the various forest ecosystems analysed by Ciais et al. (2005). Therefore it would be too early to generalize our findings to other grassland ecosystems in Europe.

With reference to climatic changes it is noteworthy that the grassland ecosystem at Rigi Seebodenalp was ready to assimilate even during the winter whenever the snow cover had disappeared. Thus, our data

suggest that any changes in duration of the winter snow covers are expected to have a strong impact on the annual CO₂ budget, even if near-surface temperatures do not change considerably.

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